

3 Species density of phanerogams and bryophytes in Dutch forests

Abstract

In Dutch forests the species density of vascular plants ranges from 1 to 61 species per 300 m². The vascular plant species density is high in the coastal dunes, southern Limburg, river valleys, and fen areas. With the exception of southern Limburg, these areas constitute the Holocene part of The Netherlands. Low species densities occur in the sandy centre of the country. To a large extent, the areas of high species density of vascular plants follow the main river valleys.

The bryophyte species densities range from 0 to 21 species per 300 m². High bryophyte species densities occur mainly in the sandy centre and in the north-eastern part of The Netherlands. The highest species densities occur in fen woodlands and derelict coppices. Bryophyte species density is low in the coastal dunes and the very young woodlands in the recently reclaimed areas (polders).

The species density contour maps of vascular plants and bryophytes in The Netherlands have little in common.

Keywords: biodiversity, species density, phanerogams, bryophytes, forests

Introduction

Biodiversity is often studied in relation to land area or some environmental factors (MacArthur 1972; Wright 1983; Peterken & Game 1984; Currie 1991; Wohlgemuth 1993). Few studies deal with the geographical pattern of biodiversity per unit area (Peet 1978; Van der Meijden et al. 1989; Currie 1991; Prendergast et al. 1993a, 1993b; Gaston & David 1994). The geographical pattern of biodiversity could be useful in evaluating environmental planning or designs for nature reserves. The ambiguous concept of biodiversity is unsuitable for direct measurement. Therefore, we use species density instead, which is an important dimension of biodiversity.

Species density covers the number of species of a particular taxonomic group per unit area. Here, the species density is the number of species per 300 m². Species density is often called species richness or species abundance (Peet 1974). It has the advantage of being easily measured. The taxonomic groups are vascular plants and bryophytes (mosses and liverworts). The vascular plant species include trees, shrubs, and herbs.

We aim at showing the geographical pattern in species density of vascular plants and bryophytes in Dutch forests.

Material and methods

Our material originates from the data base of the Fourth Dutch Forest Inventory (Anonymus 1985; Dirkse 1987). This data base contains about 2,000 vegetation records, which were made during 1984 and 1985. These records constitute a stratified random sample of the Dutch forests, which was taken in order to estimate the importance of these forests for nature conservation purposes. The stratification parameters include main tree species, stand age, and site class.

The circular sample plots measure 300 m². The centre of the plots was made permanent by a buried coil sealed in a durable waterproof coating. The subterranean coils allow exact relocation and reconstruction of the sample plots, which is a prerequisite for monitoring (Dawkins 1970; Goldsmith 1991).

The number of vegetation records per 2,500 ha ranges from 1 to more than 7 (Fig. 3.1). In The Netherlands, squares of 2,500 ha are being used as units in species distribution maps (Van der Meijden et al. 1989). The mean number amounts to 3. Higher numbers of vegetation records per 2,500 ha (to more than 7) are mostly situated in the Pleistocene parts of The Netherlands, particularly the Utrechtse Heuvelrug and the Veluwe, because most Dutch forests grow there. Lower numbers (1 and 2) occur mainly in the poorly forested Holocene part of the country. Vegetation records are absent from most Frisian Islands in the Wadden Sea and from SW Friesland (Gaasterland).

More methodological details of the Fourth Dutch Forest Inventory may be found in the reports of the Dutch State Forest Service (Anonymus 1988), Dirkse (1987), and Ritskes & Daamen (1987).

The contour maps of the species densities in Dutch forests have been produced in two steps by the computer program SURFER (Anonymus 1990). First, SURFER converted the randomly located observation points to a grid (gridding). We used inverse distance gridding:

$$Z = \frac{\sum_{i=1}^n (Z_i^1 / (d_i)^m)}{\sum_{i=1}^n (1 / (d_i)^m)}$$

Z is the value of the point being estimated,
 n is the maximum of neighbouring (nearest) points,
 Z_i is the value of a point neighbouring Z,
 d_i is the distance from Z_i to Z,
 m is a weighting power.

The greater the distance to a point (d_i), the less it contributes to the value of the point being estimated (Z). We used inverse distance gridding with a weighting power (m) of 2 and a maximum of 3 nearest points (n).

In addition, we chose a search radius of 10 km and a grid size of 5 km². The search radius limits the search area for nearest points. The grid size sets the distance between the points estimated.

The second step produces the contour lines by interpolating the values at the grid points. This step merely generalizes and polishes the gridded values.

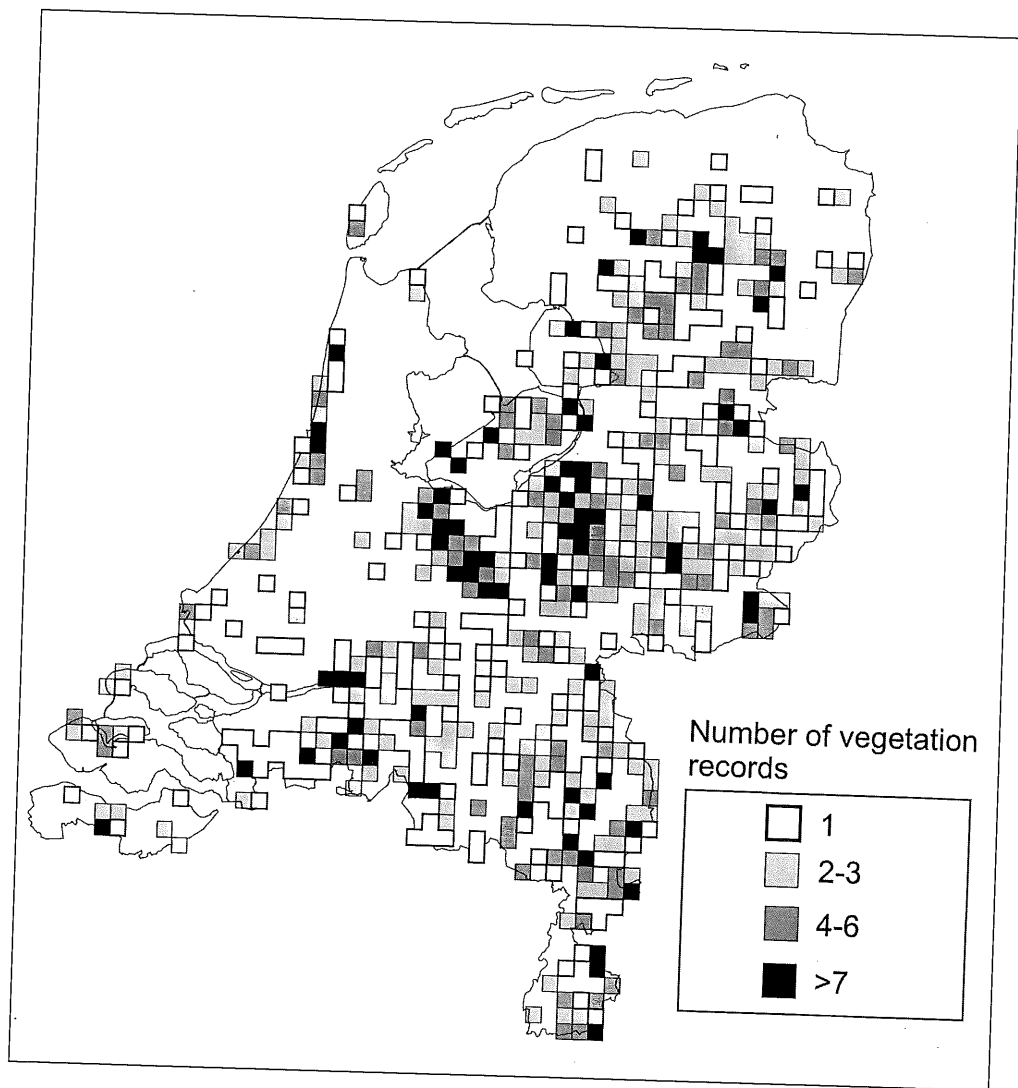


Figure 3.1
Number of forest vegetation records per 5 km² (2,500 ha).

Site description

Forests occupy about 10% of the Dutch land area (Anonymus 1988). Nearly all forests have been planted. Most plantations are in former heathlands. In a nutshell, the Dutch forest is a mixed forest of Common oak (*Quercus robur*), birch (*Betula* spp.) and Scots pine (*Pinus sylvestris*), with a shrub layer of Mountain ash (*Sorbus aucuparia*) and brambles (*Rubus* spp.) above a field layer of grasses and ferns (Dirkse & Thalen 1987). A classification resulting from the application of TWINSpan (Hill 1979) provides a more detailed picture of the Dutch forests.

Two main forest vegetation types occur (Dirkse & Thalen 1987):

1 a type with *Deschampsia flexuosa* and *Molinia caerulea*. 82% of the Dutch forest vegetation belongs to this *Deschampsia flexuosa* type. It is represented mainly under Scots pine (*Pinus sylvestris*), Common oak (*Quercus robur*), and birch (*Betula* spp.);

2 a type with *Urtica dioica*, *Galium aparine*, and *Poa trivialis*. 18% of the Dutch forest vegetation belongs to the *Urtica dioica* type. A field layer of this type is developed under deciduous trees: mainly willow (*Salix* spp.), poplar (mainly *Populus x canadensis*), and ash (*Fraxinus excelsior*).

The forest types inhabit separate areas (Fig. 3.2): the *Deschampsia flexuosa*-type is almost restricted to areas with Pleistocene sandy sediments, mostly acid and poor in nutrients. The *Urtica dioica* type is almost confined to areas with Holocene sediments (clay, peat, and sand), mostly neutral or basic and much richer in nutrients. In The Netherlands, the geographical distribution of the Pleistocene and the Holocene constitutes an important ecological structure. The distribution of the two main forest types roughly reflects this structure. The main forest types may be further divided into subunits (Van der Werf 1991; Dirkse 1993).

Results

Vascular plant species density

The number of vascular plant species in the sample plots of 300 m² (species density) ranges from 1 to 61 (Fig. 3.3). The mean species density is about 12.

The lowest species densities have been recorded in young conifer stands, particularly Douglas fir (*Pseudotsuga menziesii*) and Norway spruce (*Picea abies*) whereas the highest species densities occur in stands of deciduous trees. Species densities below 5 are rare in deciduous forests and densities above 30 have been recorded frequently (Dirkse 1987).

Two extremely high species densities merit attention: 54 and 61. The highest species density (61) was recorded in a stand of poplars, next to a waste dump. The stand contained many annual weeds and ruderals. The second highest species density (54) was found in a clearance area of a poplar stand crowded with annual weeds, grasses and ruderals. Obviously, in The Netherlands extremely high species densities do not necessarily indicate forests deserving to rank high on protection priority lists. This may also apply to other densely populated areas in Europe (Rebele 1994).

In making the contour maps, it appeared that the two highest species densities (54, 61) affected the contours too strongly. Therefore, these two extremes were omitted.

The contour map for vascular plants (Fig. 3.4) shows a lowest density contour of 10 species per 300 m², a contour interval of 10 species, and a highest contour of 30 species per 300 m².

The lowest species densities amount to 10 vascular plant species. These low densities are confined to the central and eastern parts of The Netherlands. They extend over a vast area in the sandy central part of The Netherlands (Veluwe, Utrechtse Heuvelrug, Salland), NE Drenthe, and the eastern part of the province of

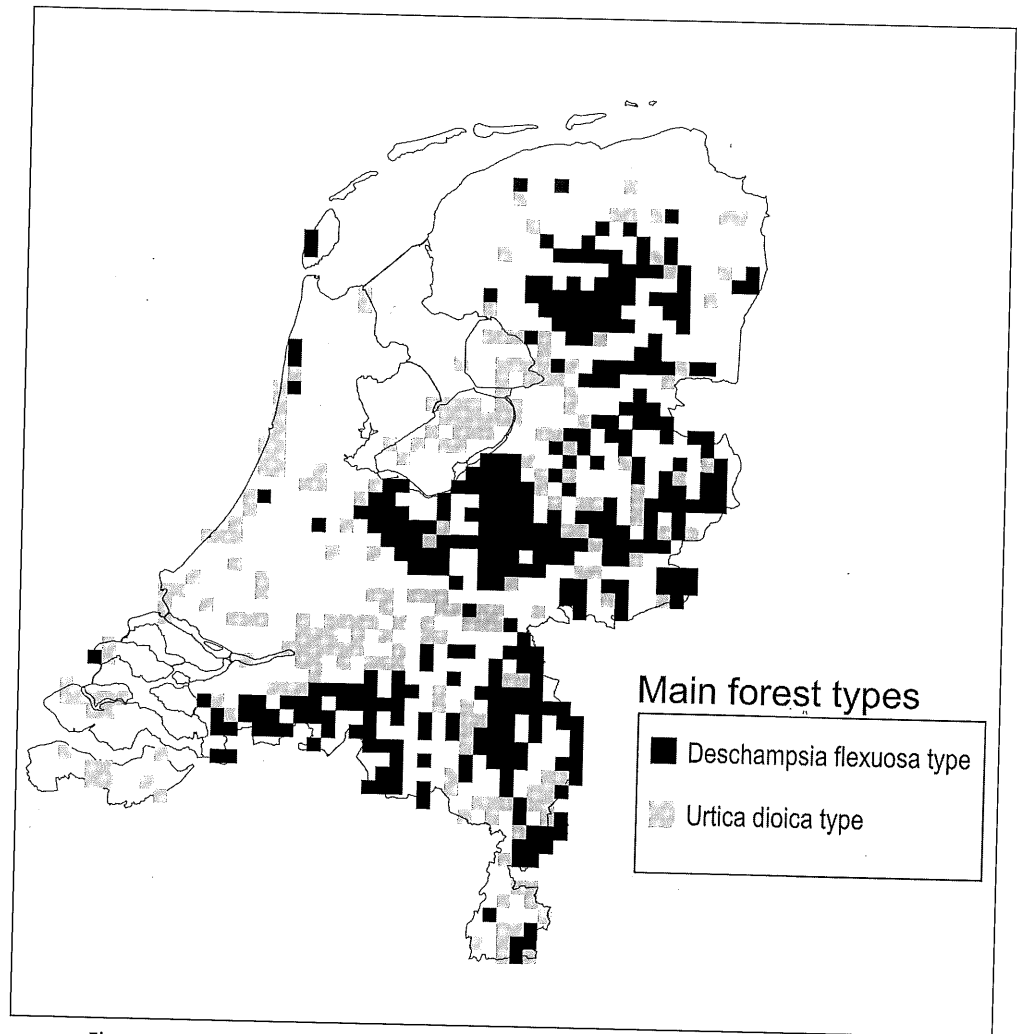


Figure 3.2
Geographical distribution of the two main forest types in The Netherlands. The *Deschampsia flexuosa* type indicates the elevated sandy part. The *Urtica dioica* type indicates the low-lying part.

Noord-Brabant. In these areas, forests of Scots pine (*Pinus sylvestris*) or Corsican pine (*Pinus nigra*) prevail (Anonymus 1985).

The areas of low species density are surrounded by extensive areas where species density ranges from 11 to 20. Far the greater part of Dutch forests falls in this category of species density.

The higher contour line encloses species densities of 20 vascular plant species or more. High species densities seem to be scattered over The Netherlands. However, three large high density areas emerge: the dune area, southern Limburg, and the middle part of the river valleys of Rhine and Meuse.

The coastal dunes constitute a long series of high species density forests. In the southern part the species density rises above 30 species. Residences and some

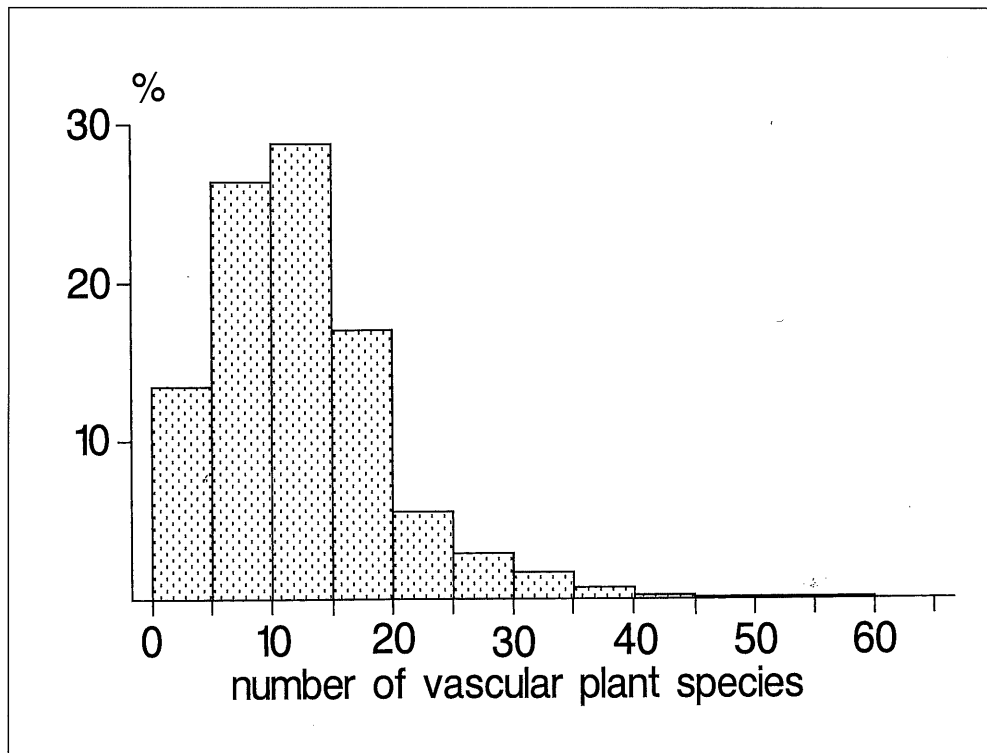


Figure 3.3
 Frequency of the number of vascular plant species present in 1914 sampling plots of the Fourth Dutch Forest Inventory.

derelict coppices form an important part of the high species density forests in the dunes.

The southeastern extremity of The Netherlands (southern Limburg) is well known for its fine woodlands which harbour many rare species (Van der Meijden et al. 1990). It is the largest limestone area in The Netherlands. The species densities are well above 20 and even reach beyond 30.

The river valleys constitute a furcated pattern of small, isolated high species density forests. This fragmented forest chain starts in the southwest, and it furcates east of the Biesbosch, a former freshwater tidal area with extensive derelict willow coppices. The southern branch extends eastwards for more than 60 kilometers. This branch is constituted by small forests which are of very different nature: coppices (whether derelict or still cut), residences, forest parks, and decoy forests. The northern branch follows the narrow valley of the river Vecht where large residences, ash coppices, and extensive fen woodlands occur.

More isolated areas of high species density occur in the northeastern part of The Netherlands. Among these areas are fen woodlands, and some ancient woodlands.

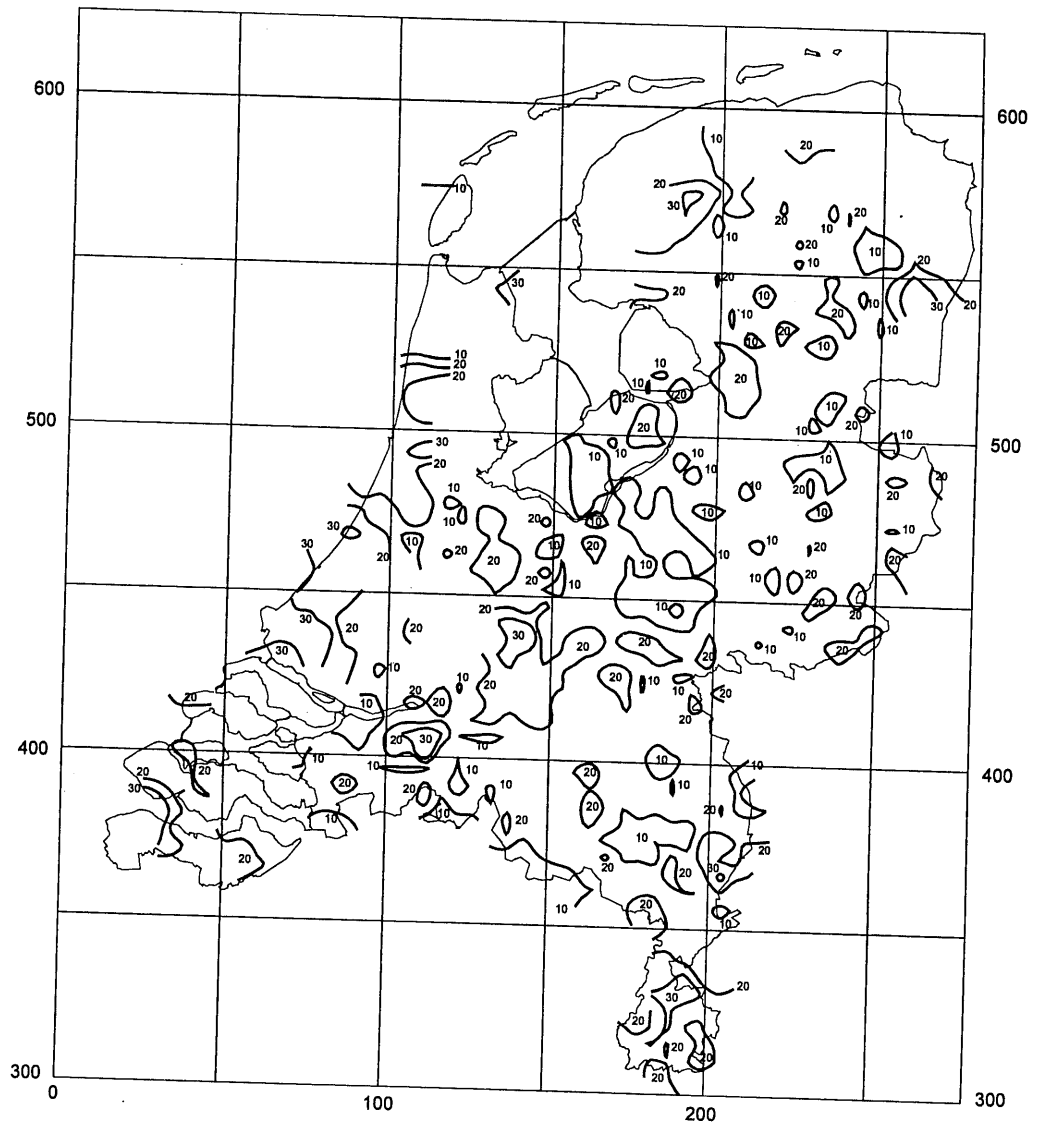


Figure 3.4
Contours of vascular plant species density (species richness) in Dutch forests. Contour interval = 10 species.

Some areas of intermediate species density (15-20 species per 300 m²) connect the high density patches such as to form branches that follow the valleys of the main rivers and their largest tributaries.

A statistical analysis (generalized linear model: McCullagh & Nelder 1989) revealed the main factors that significantly ($p < 0.01$) affect the species density of vascular plants in Dutch forests. For this analysis we used the vegetation records ($n=1912$) and a corresponding set of environmental variables (Dirkse 1987). The



significance was obtained from the deviance ratio (McCullagh & Nelder 1989).

The main density affecting factors include (in order of decreasing significance): vegetation type, herb layer cover, moss cover, soil humus content, main tree species, and tree cover.

Soil class, forest type according to age, groundwater level, and thickness of the litter layer did not significantly affect the species density. Forest type according to age includes three age classes: <1800; 1800-1900; >1900.

Bryophytes

The bryophytes include species from four forest habitats: the forest floor, decaying wood, tree bases, and tree stems (Dirkse 1987). The bryophyte species density ranges from 0 to 23 species per 300 m² (Fig. 3.5). The mean species density is about 7. The highest bryophyte species densities occur in coniferous forests (Dirkse 1987).

The lowest contour on the contour map (Fig. 3.6) indicates a species density of 4 bryophyte species or less, the contour interval is 4 species, and the highest contour indicates 16 species or more.

Low bryophyte species densities (up to 4 species) occur scattered over The Netherlands, the largest areas are situated in the coastal dunes, the polder area slightly above the middle of the country, and the northeast. Many small patches occur in the east and southeast of The Netherlands.

Large parts of the east and southeast have low to intermediate species densities (4-8 species).

In the central part of The Netherlands, the northeast, and part of the southwest high bryophyte species densities occur (8-16 species). In the centre these high densities are in the elevated areas of Pleistocene sands. For example, in the centre of The Netherlands regions of high bryophyte species density (12-16 species) mark the southern contours of the Utrechtse Heuvelrug and the Veluwe (Veluwezoom).

In the low-lying part of The Netherlands, high bryophyte species densities occur in willow and ash coppice, decoy forests, and fen woodland. The highest bryophyte species densities, 16 bryophyte species or more per 300 m², occur along the lower reaches of the river Vecht. Residences, derelict ash coppices and fen woodlands (Vechtplassen) contribute to the high bryophyte species density in this area. Spots of very high bryophyte species density (16 species) also occur in species rich areas (8-12 species) in the northeast and the southwest of The Netherlands.

High bryophyte species densities (12 or more) are very rare in the coastal dunes and the southeast of The Netherlands.

Like the numbers of vascular plant species, we analysed statistically the factors that affect the number of bryophyte species. Factors significantly ($p < 0.01$) affecting the bryophyte species density include (in order of decreasing significance): vegetation type, moss cover, forest type according to age, main tree species, cover of herb layer, and the thickness of the litter layer.

Groundwater level, soil class, and tree cover do not significantly affect the bryophyte species density.

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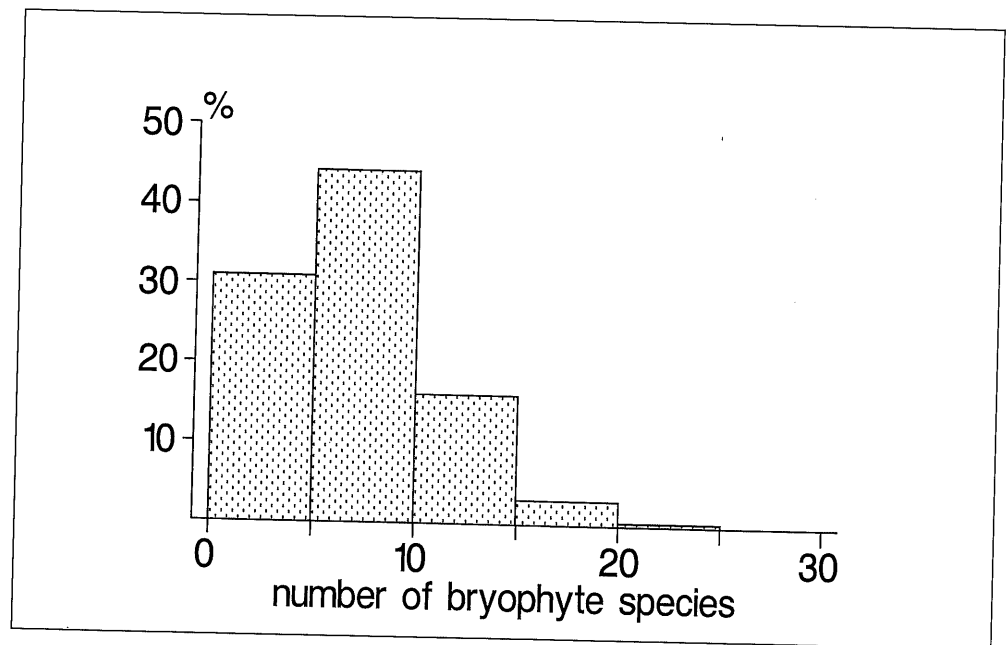


Figure 3.5
Frequency of the number of bryophyte species present in 1914 sampling plots of the Fourth Dutch Forest Inventory.

Recent changes

In 1994 we used about 100 permanently marked plots of the Fourth Dutch Forest Inventory in studying recent vegetation changes in conifer and broad-leaved forests in Twente, situated in the eastern part of The Netherlands, bordering Germany (Samsen 1995; Dirkse & Samsen 1998). The vegetation of these stands also had been recorded in 1984 or 1985. Due to the permanent plot markings we could exactly relocate the plots which had been laid out ten years earlier.

In 1984, the mean species density of vascular plants was 11.5. In 1994, the species density was 16.3. So, in ten years, the mean species density of vascular plants had increased by 4.8 species.

In the same period, the mean bryophyte species density had increased by 5.8 species: in 1984 an average of 7.1 bryophyte species per 300 m² was found while in 1994 the average density of bryophyte species was 12.9.

These results agree with those of Van Dobben et al. (1994), who re-recorded the vegetation of 177 stands of Scots pine that were part of the stratified random sample of the Fourth Dutch Forest Inventory. In 1993, the mean species density of vascular plants in these pine forests had increased significantly by 3.2 species; the mean bryophyte species density had increased by 4.7 species.

Both Van Dobben et al. (1994) and Samsen (1995) convincingly support the conclusion that in Dutch forests the species densities of vascular plants and bryophytes are increasing. This applies particularly to Scots pine forests.

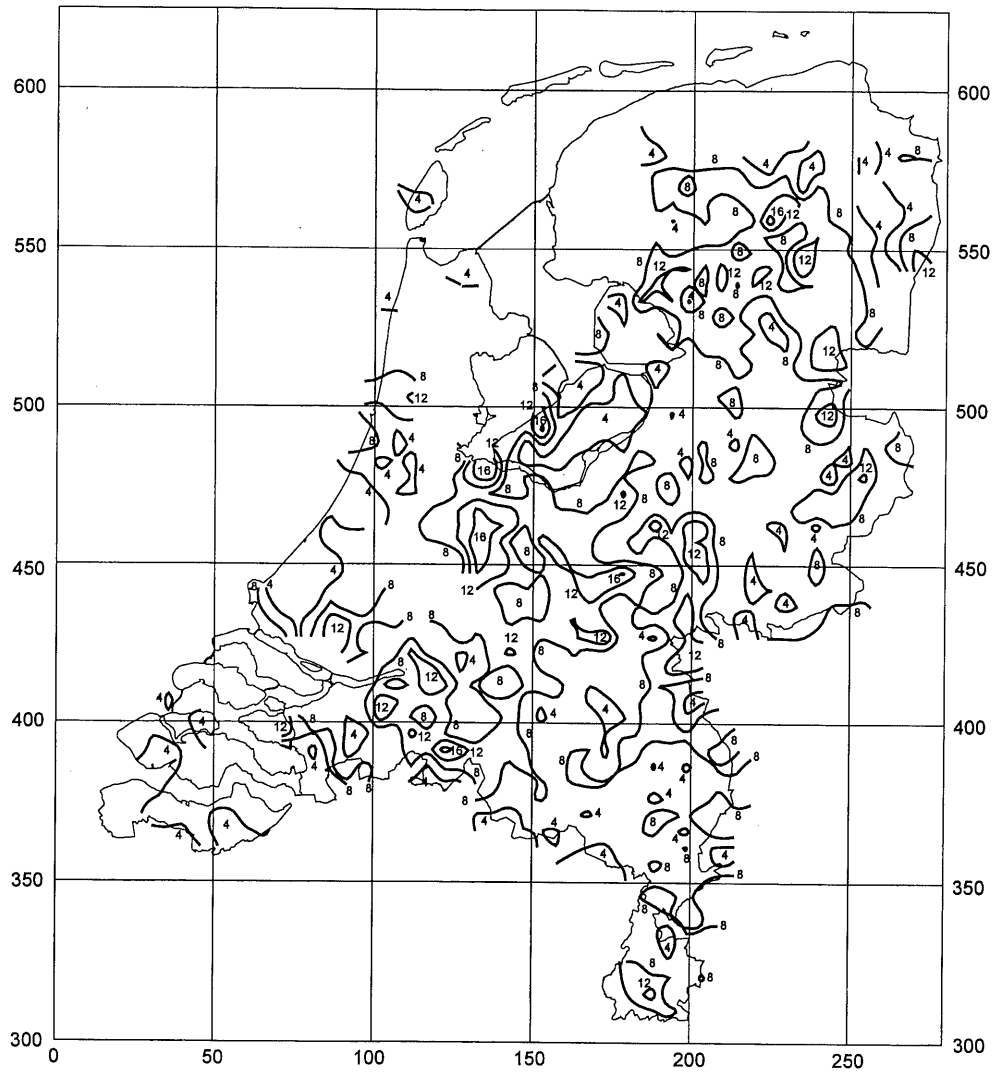


Figure 3.6
 Contours of bryophyte species density (species richness) in Dutch forests. Contour interval = 4 species.

Discussion and conclusion

The species density contour maps of vascular plants and bryophytes in The Netherlands have little in common. To some extent, the maps even show opposite patterns of species density. For example, the high sandy soils in the centre and the northeast of The Netherlands have low densities of vascular plant species, but high densities of bryophyte species. The opposite applies to the coastal dune area, where the density of vascular plant species is high but the bryophyte species density is low.

Another example of exclusive patterns in species densities of vascular plants and bryophytes can be found in the valleys of the main rivers, where the density of vascular plant species is high, whereas the density of bryophyte species is usually low. Both in the dunes and the river valleys, the low bryophyte species density is probably caused by a high cover of the herb layer. Only in some fen woodlands, some derelict coppices and some decoy forests, the species densities of vascular plants and bryophytes are comparably high.

In Dutch forests, diversity hot spots for vascular plants and mosses seldom coincide. This conclusion supports the one of Prendergast et al. (1993a), who examined the extent to which species-rich areas for five higher taxa (Butterflies, Dragonflies, Liverworts, Aquatic plants, and Breeding birds) coincide. They used British distribution maps with a grid of 10 km squares and found that 'species-rich areas ("hot spots") frequently do not coincide for different taxa'. Gaston and David (1994) used species distribution maps of twelve higher taxa in examining the coincidence of hot spots of species richness across Europe. The distribution maps used had a grid of (approximately) 152,000 km². At this scale, a significant coincidence of hotspots was found. Apparently, incongruence in local diversity of higher taxa needs not preclude congruence in diversity of these taxa at a 'megascale'.

Over the last ten years, the species density of vascular plants and bryophytes in Dutch forests have increased considerably. How this increase affects the geographical pattern of species density, remains to be explored. The increase in a region may deviate from a national trend.

The contour map of vascular plant species densities (Fig. 3.4) applies to forests, which occupy only 10% of the land area. Most of the mapped area consists of land use categories other than forests: agricultural land (71%), urban or industrial areas (13%), and non forest nature (3%). Therefore, the forest species density map (Fig. 3.4) does not necessarily reflect the general pattern of vascular plant species density. Van der Meijden et al. (1989) provided a general view of the vascular plant species densities in The Netherlands, based on species counts in 1543 5x5 km² squares. The highest species densities amount to 451 species or more per square. These occur in the dune area, the eastern and southern borders of the Utrechtse Heuvelrug and the Veluwe, some parts of the large river valleys, the central part of Noord-Brabant, and a large part of southern Limburg. Areas with low species densities are the Veluwe, a large part of Drenthe, and the eastern part of Noord-Brabant. The contours of the forest species density reflect this pattern to a large extent. This could imply that the general species density of a landscape is reflected by the species densities of its forests.

In order to get an idea of the reliability of the contour maps, we raised the number of observation points to be taken into account by the gridding facility of the computer program SURFER. Setting this number to 12 results in a less patchy contour map. The raised number of observation points caused the grid values to be closer to the mean. As a consequence, patches of extreme densities got smaller or disappeared. The position of the contour lines near the average species density remained almost unchanged. The levelling effect depends on the original number of observation points within the search area, relative to the optional number for gridding. Contour lines in regions with few observation points within the search area remain unchanged by a higher number of observation points desired for gridding.

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For two reasons, randomly chosen points are not maximally suited for making contour maps with a Geographical Information System (Jongman 1990). First, some area may by chance be unrepresented in a random sample. For example, the vegetation data of the Fourth Dutch Forest Inventory do not have records from the West Frisian Islands other than Texel. Moreover, SW Friesland (Gaasterland) is unrepresented in the data.

The second reason is more technical. Prior to producing a contour map, the irregularly spaced points have to be converted into a grid (gridding), which means not only a lot of displacement of observation localities, but also a lot of interpolation. Both the displacement and the extra interpolation cause inaccuracy and artefacts. Moreover, the gridding procedure requires many personal choices between options provided by the computer program used. Therefore, if vegetation observations are going to be made for contour mapping, the observation points should better not be randomly placed but regularly.

Obviously, regularly placed observation points need no gridding before contour mapping. Therefore, these points serve better in making reliable contour maps.

The contour maps that we produced make sense, but contour maps based on regularly spaced observation points would probably do so more accurately.

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References

- Anonymous 1990. Reference Manual SURFER Version 4. Golden Software, Golden, Colorado.
- Anonymus 1985. De Nederlandse Bosstatistiek. Deel 1: de oppervlakte bos 1980-1983. Staatsuitgeverij/CBS-publikaties, 's-Gravenhage.
- Anonymus 1988. Instructie veldopname Vierde Nederlandse Bosstatistiek. Rapport 1988-22. Staatsbosbeheer, Utrecht.
- Anonymus 1988-1990. Luchtkwaliteit - jaarverslagen 1987-1989. Rapporten Rijksinstituut voor Volksgezondheid en Milieuhygiëne, Bilthoven.
- Ashby, E. 1936. Statistical ecology. *Botanical Review* 2: 221-235.
- Asman, W.A.H. & H.A. van Jaarsveld 1990. A variable-resolution statistical transport model applied for ammonia and ammonium. Report 228471007. Rijksinstituut voor Volksgezondheid en Milieuhygiëne, Bilthoven.
- Austin, M.P. & P.C. Heyligers 1989. Vegetation survey design for conservation: gradsect sampling of forests in Northeastern New South Wales. *Biological Conservation* 50: 13-32.
- Barkman, J.J., H. Doing Kraft, C.G. van Leeuwen & V. Westhoff 1958. Enige opmerkingen over de terminologie in de vegetatiekunde. *Correspondentieblad ten dienste van de floristiek en het vegetatie-onderzoek van Nederland*, verzorgd door de Afdeling Nederland van het Rijksherbarium, in samenwerking met het I.V.O.N. 8: 87-93.
- Barkman, J.J., J. Moravec & S. Rauschert 1986. Code of phytosociological nomenclature. *Vegetatio* 67: 145-195.
- Beck, M.W. 1997. Inference and generality in ecology: current problems and an experimental solution. *Oikos* 78: 265-273.
- Becking, R.W. 1957. The Zürich-Montpellier school of phytosociology. *Botanical Review* 23: 411-488.
- Braun-Blanquet, J. 1928. *Pflanzensoziologie*. Springer, Berlin.
- Braun-Blanquet, J. 1951. *Pflanzensoziologie. Grundzüge der Vegetationskunde*. Springer, Wien.
- Braun-Blanquet, J. 1964. *Pflanzensoziologie*. Springer, Wien.
- Bremer, P. 1997. De ontwikkeling van de bosflora in de Flevolandse kleibossen. P. Bremer, Zwolle.
- Brunet, J., U. Falkengren-Grerup, A. Rühling & G. Tyler 1997. Regional differences in floristic change in south Swedish oakforests as related to soil chemistry and land use. *Journal of Vegetation Science* 8: 329-336.
- Bunschoten, L. 1987. Overzicht van de geschiedenis van de Nederlandse bosstatistiek. *Nederlands Bosbouw tijdschrift* 59: 76-78.
- Cohen, J. 1960. A coefficient of agreement for nominal scales. *Educational Psychology Measures* 20: 37-46.
- Currie, D.J. 1991. Energy and large-scale pattern of animal- and plant-species richness. *The American Naturalist* 137: 27-49.
- Dawkins, H.C. 1970. Techniques for long-term diagnosis and prediction in forest communities. In: E. Duffey & A.S. Watt, *The scientific management of animal and plant communities for conservation*. Blackwell, Oxford etc.
- De Vries, I.M. 1982. De invloed van luchtverontreiniging/zure neerslag op hogere planten; november 1980-november 1981. Rijksinstituut voor Natuurbeheer, Leersum.
- Dierschke, H. 1994. *Pflanzensoziologie*. Ulmer, Stuttgart.

- Dierssen, C. & B. Dierssen 1996. Vegetation Nordeuropas. Ulmer, Stuttgart.
- Dirkse, G.M. 1987. De natuur van het Nederlandse bos. RIN-rapport 87/28. Rijksinstituut voor Natuurbeheer, Leersum.
- Dirkse, G.M. 1993. Bostypen in Nederland. Wetenschappelijke Mededeling 208. Stichting Uitgeverij KNNV, Utrecht.
- Dirkse, G.M. 1994. The Fourth Dutch Forest Survey, a stratified random approach to vegetation sampling. Colloques Phytosociologiques XXIII: 401-416.
- Dirkse, G.M. 1997. Vegetatiekartering van de Schinveldse bossen en de Brunsummerheide in 1996. IBN-rapport 261. Instituut voor Bos- en Natuuronderzoek, Wageningen.
- Dirkse, G.M. & H.F. van Dobben 1989. Effects of experimental fertilization on forest undergrowth in young stands of Scots pine in Sweden. *Studies in Plant Ecology* 18: 62-64.
- Dirkse, G.M. & H.F. van Dobben 1998. Changes in the flora of Dutch forests of Scots pine (*Pinus sylvestris*) between 1984-1985 and 1993. Chapter 5 in this thesis.
- Dirkse, G.M. & G.F.P. Martakis 1992. Effects of fertilizer on bryophytes in Swedish experiments on forest fertilization. *Biological Conservation* 59: 155-161.
- Dirkse, G.M. & G.F.P. Martakis 1993. Recent changes in forest vegetation in North-West and Central Europe and some likely causes. In: M.E.A. Broekmeyer, W. Vos & H. Koop (eds.), *European Forest Reserves*. Pudoc, Wageningen; 233-247.
- Dirkse, G.M. & G.F.P. Martakis 1998. On the local validity of syntaxa. Chapter 4 in this thesis.
- Dirkse, G.M., G.F.P. Martakis & P.W. Goedhart 1998. On the ageing of syntaxa. Chapter 7 in this thesis.
- Dirkse, G.M., H.M.H. van Melick & A. Touw 1988. Checklist of Dutch bryophytes. *Lindbergia* 14: 167-175.
- Dirkse, G.M. & F.J. Samsen 1998. Changes in the forest flora in Twente (province of Overijssel, The Netherlands) between 1984-1985 and 1994. Chapter 6 in this thesis.
- Dirkse, G.M. & D.P.C. Thalen 1987. De 'natuurfunctie' van het Nederlandse bos; enkele resultaten van de Vierde Bosstatistiek. *Nederlands Bosbouw tijdschrift* 59: 116-127.
- Doing, H. 1962. Systematische Ordnung und floristische Zusammensetzung niederländischer Wald- und Gebüschgesellschaften. North Holland Publishing Company, Amsterdam.
- Doing, H. 1972. Proposals for an objectivation of phytosociological methods. In: E. van der Maarel & R. Tüxen (eds.), *Grundfragen und Methoden in der Pflanzensoziologie*. Junk, Den Haag.
- Donner, A., M. Eliasziw & N. Klar 1996. Testing the homogeneity of kappa statistics. *Biometrics* 52: 176-183.
- Efron, B. & R.J. Tibshirani 1993. *An introduction to the bootstrap*. Chapman and Hall, New York.
- Ellenberg, H. 1954. Über einige Fortschritte der kausalen Vegetationskunde. *Vegetatio* 5/6: 199-211.
- Ellenberg, H. 1956. Aufgaben und Methoden der Vegetationskunde. Einführung in die Phytologie IV, 1: 1-136. Ulmer, Stuttgart.
- Ellenberg, H. 1978. Vegetation Mitteleuropas mit den Alpen. Ulmer, Stuttgart.
- Ellenberg, H. 1985. Veränderungen der Flora Mitteleuropas unter dem Einfluss von Düngung und Emissionen. *Schweizerische Zeitschrift für das Forstwesen* 136: 19-39.
- Ellenberg, H. 1991. Zeigerwerte der Gefäßpflanzen (ohne *Rubus*). *Scripta Geobotanica* 18: 1-166.
- Ellenberg, H., H.E. Weber, R. Duell, V. Wirth, W. Werner & D. Paulissen 1992. Zeigerwerte der Gefäßpflanzen Mitteleuropas. *Scripta Geobotanica* 18: 1-258.

- Falkengren-Grerup, U. 1990. Distribution of field layer species in Swedish deciduous forests in 1929-54 and 1979-88 as related to soil pH. *Vegetatio* 86: 143-150.
- Falkengren-Grerup, U. & H. Eriksson 1990. Changes in soil, vegetation and forest yield between 1947 and 1988 in beech and oak sites of southern Sweden. *Forest Ecology and Management* 38: 37-53.
- Gaston, K.J. & R. David 1994. Hotspots across Europe. *Biodiversity Letters* 2: 108-116.
- Gauch H.G. 1983. *Multivariate analysis in community ecology*. Cambridge University Press, Cambridge etc.
- Genstat 5 Committee 1993. *Genstat 5 Release 3 reference manual*. Clarendon Press, Oxford.
- Gerhardt, K. & O. Kellner 1986. Effects of nitrogen fertilization on the field- and bottomlayer species composition in some Swedish coniferous forests. *Meddelanden från Växtbiologiska Institutionen, Uppsala*. 1986 (1): 1-47.
- Glenn-Lewin, D.C., R.K. Peet & T.T. Veblen (eds.) 1992. *Plant succession; theory and prediction*. Chapman and Hall, London etc.
- Goldsmith, F.B. 1991. *Monitoring for conservation and ecology*. Chapman and Hall, London etc.
- Goldsmith, F.B. & C.M. Harrison 1976. Description and analysis of vegetation. In: S.B. Chapman (ed.), *Methods in plant ecology*. Blackwell, Oxford etc.; 85-155.
- Goodall, D.W. 1952. Quantitative aspects of plant distribution. *Biological Reviews of the Cambridge Philosophical Society* 27: 194-245.
- Gordon, A.D. 1981. *Classification*. Chapman and Hall, New York.
- Green, R.H. 1979. *Sampling design and statistical methods for environmental biologists*. Wiley, New York.
- Greigh-Smith, P. 1983. *Quantitative plant ecology*. Blackwell, Oxford.
- Hill, M.O. 1979. TWINSPAN - a FORTRAN program for arranging multivariate data in an ordered two-way table by classification of the individuals and attributes. Cornell University, Ithaca.
- Hogeweg, P. 1976. *Topics in biological pattern analysis*. Thesis, Utrecht.
- Hogeweg, P. & W.G. Beeftink 1976. Ecomorphology of vegetation succession on recently emerged flats, a numerical analysis. In: P. Hogeweg, *Topics in biological pattern analysis*. Thesis, Utrecht; 153-199.
- Hommel, P.W.F.M., E.E.J.M. Leeters & J.G. Vrieling 1991. *Veranderingen in bodem en vegetatie van het Speulderbos; kaartvergelijking 1958-1988*. Rapport 104.1. DLO-Staring Centrum, Wageningen.
- Hull, D.L. 1964. The effect of essentialism on taxonomy - Two thousand years of stasis (I). *The British Journal for the Philosophy of Science* 15: 314-326.
- Hull, D.L. 1965. The effect of essentialism on taxonomy - Two thousand years of stasis (II). *The British Journal for the Philosophy of Science* 16: 1-18.
- Ivimey-Cook, R.B. & M.C.F. Proctor 1966. The application of association-analysis to phytosociology. *Journal of Ecology* 54: 197-192.
- Jongman, H.G. 1990. Ecological classification of the climate of the Rhine catchment. *International Journal of Biometeorology* 34: 194-203.
- Jongman, R.H.G., C.J.F. ter Braak & O.F.R. van Tongeren 1995. *Data analysis in community and landscape ecology*. Cambridge University Press, Cambridge etc.
- Kenkel, N.C. & J. Podani 1991. Plot size and estimation efficiency in plant community studies. *Journal of Vegetation Science* 2: 539-544.
- Kirby, K.J., T. Bines, A. Burn, J. Mackintosh, P. Pitkin & I. Smith 1986. Seasonal and observer

- differences in vascular plant records from British woodlands. *Journal of Ecology* 74: 123-131.
- Kop, L.G. & W. van Zeist 1960. Twente natuurhistorisch II, de bodem en de bossen. Wetenschappelijke Mededeling KNNV 37. KNNV, Hoogwoud.
- Kopecky, K. & S. Hejny 1978. Die Anwendung einer deduktiven Methode syntaxonomischer Klassifikation bei der Bearbeitung der strassenbegleitenden Pflanzengesellschaften Nordostböhmens. *Vegetatio* 36: 43-51.
- Kowarik, I. 1987. Kritische Anmerkungen zum theoretischen Konzept der potentiellen natürlichen Vegetation mit Anregungen zu einer zeitgemässen Modifikation. *Tuexenia* 7: 53-67.
- Küchler, A.W. 1967. *Vegetation mapping*. Ronald Press Company, New York.
- Küchler, A.W. 1988. The legend: organizing the map. In: A.W. Küchler & I.S. Zonneveld (eds.), *Vegetation mapping*. Kluwer, Dordrecht etc.; 121-133.
- Küchler, A.W. & I.S. Zonneveld (eds.) 1988. *Vegetation mapping*. Kluwer, Dordrecht etc.
- Kuhn, N., R. Amiet & N. Hufschmid 1987. Veränderungen in der Waldvegetation der Schweiz infolge Nährstoffanreicherungen aus der Atmosphäre. *Allgemeine Forst- und Jagdzeitung* 158: 77-158.
- Lähde, E. 1987. Changes in the forest vegetation of southern Finland. *Aquilo, Series Botanica* 25: 89-95 (in Finnish).
- Lambert, F.J.D. 1998. Het voorkomen en de uitbreiding van rankende helmblom in Oost-Brabant en Limburg. *Natuurhistorisch Maandblad* 87: 67-73.
- Landis, J.R. & G.G. Koch 1977. The measurement of observer agreement for categorical data. *Biometrics* 33: 159-174.
- Langford, A.N. & M.F. Buell 1969. Integration, Identity and Stability in the Plant Association. *Advances in Ecological Research* 6: 83-135.
- Leps, J. & V. Hadincová 1992. How reliable are your vegetation analyses? *Journal of Vegetation Science* 3: 119-124.
- Leys, H.N. 1978. Handleiding ten behoeve van vegetatiekarteringen. Wetenschappelijke Mededeling 130. KNNV, Hoogwoud.
- MacArthur, R.H. 1972. *Geographical ecology*. Harper and Row, New York.
- Matzke-Hajek, G. 1993. Die Brombeeren (*Rubus fruticosus*-Agg.) der Eifel und der Niederrheinischen Bucht. *Decheniana, Beiheft* 32. Bonn.
- McCullagh, P. & J.A. Nelder 1989. *Generalized linear models*. Chapman and Hall, London etc.
- McCune, B., J.P. Dey, J.L.E. Peck, D. Cassell, K. Heiman, S. Will-Wolf & P.N. Neitlich 1997. Repeatability of community data: species richness versus gradient scores in large-scale lichen studies. *The Bryologist* 100: 40-46.
- McLean, R.C. & W.R. Ivimey-Cook 1973. *Textbook of theoretical botany* 4. Longman, London.
- Medwecka-Kornás, A. & S. Gawronski 1990. The dieback of fir *Abies alba* Mill. and changes in the Pino-Quercetum stands in the Ojców National Park (Southern Poland). *Vegetatio* 86: 175-186.
- Meijer Drees, E. 1951. Enkele hoofdstukken uit de moderne plantensociologie en een ontwerp voor nomenclatuurregels voor plantengezelschappen. Rapport van het Bosbouwproefstation 51. Bogor.
- Meltzer, J. & V. Westhoff 1944. *Inleiding tot de plantensociologie*. Breughel, 's-Graveland.
- Monserud, R.A. & R. Leemans 1992. Comparing global vegetation maps with the kappa statistic. *Ecological Modelling* 62: 275-293.
- Moravec, J. 1981. Die Logik des pflanzensoziologischen Klassifikationssystems. In: H.

- 74: Dierschke (ed.), *Syntaxonomie*. Cramer, Vaduz; 43-63.
- Mucina, L. 1997. Classification of vegetation: past, present and future. *Journal of Vegetation Science* 8: 751-760.
- Mueller-Dombois, D. & H. Ellenberg 1974. *Aims and methods of vegetation ecology*. Wiley, New York.
- Nelder, V.J., D.C. Crossley & M. Cofinas 1995. Using geographic information systems (GIS) to determine the adequacy of sampling in vegetation surveys. *Biological Conservation* 75: 1-17.
- 7: Nilsson, I.N. & S.G. Nilsson 1985. Experimental estimates of census efficiency and pseudo-turnover on islands: error trend and between-observer variation when recording vascular plants. *Journal of Ecology* 73: 65-70.
- Oberdorfer, E. 1968. Assoziation, Gebietsassoziation, geographische Rasse. In: R. Tüxen (ed.), *Pflanzensoziologische Systematik*. Junk, Den Haag.
- Oberdorfer, E. 1972. (short communication on p. 290). In: R. Tüxen (ed.), *Grundfragen und Methoden in der Pflanzensoziologie*. Junk, Den Haag.
- Oberdorfer, E. 1977. *Süddeutsche Pflanzengesellschaften*. I. Fischer, Jena.
- Oksanen, J. & P.R. Minchin 1997. Instability of ordination results under changes in input data order: explanations and remedies. *Journal of Vegetation Science* 8: 447-454.
- Palmer, M.W. & P.S. White 1994. On the existence of ecological communities. *Journal of Vegetation Science* 5: 279-282.
- a. Payne, R.W. et al. ('Genstat 5 Committee') 1987. *GENSTAT 5 Reference Manual*. Clarendon Press, Oxford.
- ion. Peet, R.K. 1974. The measurement of species diversity. *Annual Review of Ecology and Systematics* 5: 285-307.
- Peet, R.K. 1978. Forest vegetation of the Colorado Front Range: patterns of species diversity. *Vegetatio* 37: 65-78.
- Persson, S. 1980. Succession in a South Swedish deciduous wood: a numerical approach. *Vegetatio* 43: 103-122.
- Peterken, G.F. & M. Game. 1984. Historical factors affecting the number and distribution of vascular plant species in the woodlands of central Lincolnshire. *Journal of Ecology* 72: 155-182.
- etc. Podani, J. 1997. On the sensitivity of ordination and classification methods to variation in the input order of data. *Journal of Vegetation Science* 8: 153-156.
- Poore, M.E.D. 1955. The use of phytosociological methods in ecological investigations I. The Braun-Blanquet system. *Journal of Ecology* 43: 226-244.
- lon. Poore, M.E.D. 1956. The use of phytosociological methods in ecological investigations IV. General discussion of phytosociological problems. *Journal of Ecology* 44: 28-50.
- ges Prendergast, J.R., R.M. Quinn, J.H. Lawton, B.C. Eversham & D.W. Gibbons 1993a. Rare species, the coincidence of diversity hotspots and conservation strategies. *Nature* 365: 335-337.
- Prendergast, J.R., S.N. Wood, J.H. Lawton & B.C. Eversham 1993b. Correcting for variation in recording effort in analysis of diversity hotspots. *Biodiversity Letters* 1: 39-53.
- Purr, S.H. 1952. *Forest Inventory*. Ronald Press Company, New York.
- Rebele, F. 1994. Urban ecology and special features of urban ecosystems. *Global Ecology and Biogeography Letters* 4: 171-187.
- Ritskes, T.M. & W.P. Daamen 1987. Doelstelling en uitvoering Vierde Bosstatistiek. *Nederlands Bosbouw tijdschrift* 59: 70-83.

- Rost-Siebert, K. & G. Jahn 1988. Veränderungen der Waldbodenvegetation während der letzten Jahrzehnte - Eignung zur Bioindikation von Immissionswirkungen? *Forst und Holz* 43: 75-81.
- Samsen, F.J. 1995. Verandering van de ondergroei van bossen in Twente tussen 1984 en 1994. Verslag 395. Katholieke Universiteit Nijmegen, Vakgroep Aquatische Oecologie en Milieubiologie.
- Schaminée, J.H.J., A.H.F. Stortelder & V. Westhoff 1991. De identificatie en classificatie van plantensociologisch onverzadigde gemeenschappen. *Stratiotes* 2: 42-52.
- Schaminée, J.H.J., A.H.F. Stortelder & V. Westhoff 1995. De vegetatie van Nederland. Deel 1: Inleiding tot de plantensociologie - grondslagen, methoden en toepassingen. *Opulus Press*, Uppsala, Leiden.
- Smartt, P.F.M. & J.E.A. Grainger 1974. Sampling for vegetation survey: some aspects of the behaviour of unrestricted, restricted, and stratified techniques. *Journal of Biogeography* 1: 193-206.
- Stace, C. 1997. *New flora of the British Isles*. Cambridge University Press, Cambridge.
- Stortelder, A.H.F., J.H.J. Schaminée & I.S. Zonneveld 1995. Vegetatiekartering. In: J.H.J. Schaminée, A.H.F. Stortelder & V. Westhoff (eds.), *De vegetatie van Nederland* 1. *Opulus Press*, Uppsala, Leiden; 211-224.
- Sýkora, K.V., L.J. de Nijs & T.A.H.M. Pelsma 1993. Plantengemeenschappen van Nederlandse wegbermen. Stichting Uitgeverij KNNV, Utrecht.
- Tansley, A.G. & T.F. Chippis (eds.) 1926. *Aims and methods in the study of vegetation*. British Empire Vegetation Commission & Crown Agents for Colonies, London.
- Tausch, R.J., D.A. Charlet, D.A. Weixelman & D.C. Zamudio 1995. Patterns of ordination and classification instability resulting from changes in input data order. *Journal of Vegetation Science* 6: 897-902.
- Ter Braak, C.A.J. 1987. CANOCO - a FORTRAN program for canonical community ordination by [partial] [detrended] [canonical] correspondence analysis and redundancy analysis (version 2.1). + update notes 1990. TNO Institute of Applied Computer Science, statistics department, Wageningen.
- Thimonier, A., J.L. Dupouey & J. Timbal 1992. Floristic changes in the herb-layer vegetation of a deciduous forest in the Lorraine Plain under the influence of atmospheric deposition. *Forest Ecology and Management* 55: 149-167.
- Touw, A. & W.V. Rubers 1989. *De Nederlandse bladmossen*. Stichting Uitgeverij KNNV, Utrecht.
- Tüxen, R. 1937. Die Pflanzengesellschaften Nordwestdeutschlands. *Mitteilungen der floristisch-soziologischen Arbeitsgemeinschaft Niedersachsen* 3: 1-170.
- Tüxen, R. 1956. Die heutige potentielle natürliche Vegetation als Gegenstand der Vegetationskartierung. *Angewandte Pflanzensoziologie* 13: 1-42.
- Tyler, G. 1987. Probable effects of soil acidification and nitrogen deposition on the floristic composition of oak (*Quercus robur* L.) forest. *Flora (Jena)* 179: 165-170.
- Van Breemen, N. & H.F.G. van Dijk 1988. Ecosystem effects of atmospheric deposition of nitrogen in The Netherlands. *Environmental Pollution* 54: 249-274.
- Van de Beek, A. 1978. Bramen in Zuid-Limburg. *Gorteria* 9: 80-88.
- Van den Burg, J. 1996. De betekenis van bodem en klimaat voor het Nederlandse bos. *IBN Scientific Contributions* 4. Instituut voor Bos- en Natuuronderzoek, Wageningen.
- Van den Wijngaard, J.K.R. 1980. De bossen van Nederland. *Nederlands Bosbouw tijdschrift* 52: 56-63.

- Van der Meijden, R., C.L. Plate & E.J. Weeda 1989. Atlas van de Nederlandse flora 3. Minder zeldzame en algemene soorten. Rijksherbarium, Leiden.
- Van der Meijden, R., E.J. Weeda, W.J. Holverda & P.H. Hovenkamp 1990. Heukels' Flora van Nederland. Wolters-Noordhoff, Groningen.
- Van der Meijden, R., E.J. Weeda, W.J. Holverda & P.H. Hovenkamp 1993. Heukels' Flora van Nederland. Wolters-Noordhoff, Groningen.
- Van der Meijden, R. 1996. Heukels' Flora van Nederland. Wolters-Noordhoff, Groningen.
- Van der Werf, S. 1991. Bosgemeenschappen. Pudoc, Wageningen.
- Van Dobben, H.F. 1993. Vegetation as a monitor for deposition of nitrogen and acidity. Thesis, Utrecht.
- Van Dobben, H.F., M.J.M.R. Vocks, E. Jansen & G.M. Dirkse 1994. Veranderingen in de ondergroei van het Nederlandse dennenbos over de periode 1985-1993. IBN-rapport 085. Instituut voor Bos- en Natuuronderzoek, Wageningen.
- Van Egmond, N.D., O. Tissing, D. Onderdelinden & C. Bartels 1978. Quantitative evaluation of mesoscale air pollution transport. Atmospheric Environment 12: 2279-2287.
- Wagner, H. 1968. Prinzipienfragen der Vegetationssystematik. In: R. Tüxen (ed.), Pflanzensoziologische Systematik. Junk, Den Haag.
- Weber, H.E. 1981. Kritische Gattungen als Problem für die Syntaxonomie der Rhamno-Prunetea in Mitteleuropa. In: H. Dierschke (ed.), Syntaxonomie. Cramer, Vaduz; 477-496.
- Weber, H.E. 1985. Rubi Westphalici - Die Brombeeren Westfalens und des Raumes Osnabrück (Rubus L., Subgenus Rubus). Abhandlungen aus dem Westfälischen Museum für Naturkunde 47, 3: 1-452.
- Weber, H.E. 1995. Gustav Hegi illustrierte Flora von Mitteleuropa, Band IV. Teil 2A Spermatophyta: Angiospermae: Dicotyledones 2,2. Blackwell, Berlin.
- Werger, M.J.A. 1973. Phytosociology of the upper Orange River Valley, South Africa; a syntaxonomical study. Thesis, Nijmegen.
- Werger, M.J.A. 1974. The place of the Zürich-Montpellier method in vegetation science. Folia Geobotanica et Phytotaxonomica 9: 99-109.
- Westfall, R.H. 1992. Objectivity in stratification, sampling and classification of vegetation. Thesis, University of Pretoria.
- Westhoff, V. 1979. Phytosociology in The Netherlands: history, present state, future. In: M.J.A. Werger (ed.), The study of vegetation. Junk, Den Haag; 81-121.
- Westhoff, V. & A.J. den Held 1975. Plantengemeenschappen in Nederland. Thieme, Zutphen.
- Westhoff, V. & E. van der Maarel 1973. The Braun-Blanquet approach. In: R.H. Whittaker (ed.), Handbook of Vegetation Science V: Ordination and classification of vegetation. Junk, Den Haag; 619-726.
- Whittaker, R.H. 1962. Classification of natural communities. Botanical Review 28: 1-214.
- Wittig, R., H.J. Ballach & C.J. Brandt 1985. Increase of acid indicators in the herb layer of Millet Grass-Beech Forest of the Westphalian Bight. Angewandte Botanik 59: 219-232.
- Wohlgemuth, T. 1993. Der Verbreitungsatlas der Farn- und Blütenpflanzen der Schweiz (Welten und Sutter 1982) auf EDV: Die Artenzahlen und ihre Abhängigkeit von verschiedenen Faktoren. Botanica Helvetica 103: 55-71.
- Wright, D.H. 1983. Species-energy theory: an extension of species-area theory. Oikos 41: 496-506.
- Zonneveld, I.S. 1988. Survey approaches. In: A.W. Küchler & I.S. Zonneveld (eds.), Vegetation mapping. Kluwer, Dordrecht etc.; 203-207.